The Conditionability, Filterability, Settleability and Solids Content of Sludges 1984 (A compendium of Methods and Tests)

Methods for the Examination of Waters and Associated Materials

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This Compendium contains the following Methods and Tests:

 Determination of the Suspended Solids Content and Sewage of Waterworks Sludges and Related Solids; 2 methods (the second in two variants). Determination of the Filterability of a Sludge using the Capillary Suction Time (CST) Method. Determination of the Conditionability of Sewage and Waterworks Sludges. Methods for the Initial Selection of Chemicals for Use as Settlement Aids in the Treatment of Wastewaters; 2 methods. Determination of the Thickening Characteristics of Sludges using a Low-Speed Centrifuge Method. Determination of the Settleability of Activated Sludge expressed as the Stirred Specific Volume Index (SSVI) or Stirred Sludge Density (SSD). Measurement of the Specific Resistance to Dewatering of a Sludge, and the Sludge 	No	te: Measurement of Suspended, Settleable and Total Dissolved Solids is given in a	
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Warning to Users

The analytical procedures given in this booklet should be carried out only by competent trained persons, with adequate supervision when necessary. Local Safety Regulations must be observed. Laboratory procedures should be carried out only in properly equipped laboratories. Field operations should be conducted with due regard to possible local hazards, and portable safety equipment should be carried. Care should be taken against creating hazards whether for one's self, one's colleagues in the laboratory, outsiders or subsequently for maintenance workers. Lone working, whether in the laboratory or field, should be discouraged. Reagents of adequate purity must be used, along with properly maintained apparatus and equipment of correct specifications. Specifications for reagents, apparatus and equipment are given in manufacturers' catalogues and various published standards. If contamination is suspected, reagent purity should be checked before use.

There are numerous handbooks on first aid and laboratory safety. Among such publications are: 'Code of Practice for Chemical Laboratories' and 'Hazards in the Chemical Laboratory' issued by the Royal Society of Chemistry, London; 'Safety in Biological Laboratories' (Editors Hartree and Booth), Biochemical Society Special Publication No 5, The Biochemical Society, London, which includes biological hazards; and 'The Prevention of Laboratory Acquired Infection', Public Health Laboratory Service Monograph 6, HMSO, London.

Where the Committee have considered that a special unusual hazard exists, attention has been drawn to this in the text so that additional care must be taken beyond that which should be exercised at all times when carrying out analytical procedures. It cannot be too strongly

emphasised that prompt first aid, decontamination or administration of the correct antidote can save life; but that incorrect treatment can make matters worse. It is suggested that both supervisors and operators be familiar with emergency procedures before starting even a slightly hazardous operation, and that doctors consulted after any accident involving chemical contamination, ingestion, or inhalation, be made familiar with the chemical nature of the injury, as some chemical injuries require specialist treatment not normally encountered by most doctors. Similar warning should be given if a biological or radio-chemical injury is suspected. Some very unusual parasites, viruses and other microorganisms are occasionally encountered in samples and when sampling in the field. In the latter case, all equipment including footwear should be disinfected by appropriate methods if contamination is suspected.

The best safeguard is a thorough consideration of hazards and the consequent safety precautions and remedies well in advance. Without intending to give a complete checklist, points that experience has shown are often forgotten include: laboratory tidiness, stray radiation leaks (including ultra violet), use of correct protective clothing and goggles, removal of toxic fumes and wastes, containment in the event of breakage, access to taps, escape routes, and the accessibility of the correct and properly maintained first-aid, fire-fighting, and rescue equipment. Hazardous reagents and solutions should always be stored in plain sight and below face level. Attention should also be given to potential vapour and fire risks. If in doubt, it is safer to assume that the hazard may exist and take reasonable precautions, rather than to assume that no hazard exists until proved otherwise.

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About this series

This booklet is part of a series intended to provide both recommended methods for the determination of water quality, and in addition, short reviews of the more important analytical techniques of interest to the water and sewage industries. In the past, the Department of the Environment and its predecessors, in collaboration with various learned societies, have issued volumes of methods for the analysis of water and sewage culminating in 'Analysis of Raw, Potable and Waste Waters'. These volumes inevitably took some years to prepare, so that they were often partially out of date before they appeared in print. The present series will be published as individual methods, thus allowing for the replacement or addition of methods as quickly as possible without need of waiting for the next edition. The rate of publication will also be related to the urgency of requirement for that particular method, tentative methods being issued when necessary. The aim is to provide as complete and up to date a collection of methods and reviews as is practicable, which will, as far as possible, take into account the analytical facilities available in different parts of the Kingdom, and the quality criteria of interest to those responsible for the various aspects of the water cycle. Because both needs and equipment vary widely, where necessary, a selection of methods may be recommended for a single determinand. It will be the responsibility of the users - the senior analytical chemist, biologist, bacteriologist etc., to decide which of these methods to use for the determination in hand. Whilst the attention of the user is drawn to any special known hazards which may occur with the use of any particular method, responsibility for proper supervision and the provision of safe working conditions must remain with the user.

The preparation of this series and its continuous revision

is the responsibility of the Standing Committee of Analysts (to review Standard Methods for Quality Control of the Water Cycle). The Standing Committee of Analysts is a committee of the Department of the Environment set up in 1972. Currently it has seven Working Groups, each responsible for one section or aspect of water cycle quality analysis. They are as follows:

- 1.0 General principles of sampling and accuracy of results
- 3.0 Empirical and physical methods
- 4.0 Metals and metalloids
- 5.0 General nonmetallic substances
- 6.0 Organic impurities
- 7.0 Biological methods
- 9.0 Radiochemical methods.

The actual methods and reviews are produced by smaller panels of experts in the appropriate field, under the overall supervision of the appropriate working group and the main committee. The names of those associated with this method are listed inside the back cover. Publication of new or revised methods will be notified to the technical press, whilst a list of Methods in Print is given in the current HMSO Sectional Publication List No.5.

Whilst an effort is made to prevent errors from occurring in the published text, a few errors have been found in booklets in this series. Correction notes and minor additions to published booklets not warranting a new booklet in this series will be issued periodically as the need arises. Should an error be found affecting the operation of a method, the true sense not being obvious, or an error in the printed text be discovered prior to sale, a separate correction note will be issued for inclusion in that booklet.

L R PITTWELL Secretary

31 October 1983

Determination of the Total Solids Content (Dry Residues at 105°C) and the Loss on Ignition of Dry Residue at 550°C of Sewage and Waterworks Sludges and Related Solids

Performance Characteristics of the Methods

1.1	Parameters determined	 (a) Total solids content. (b) Loss on ignition of dry residue at 550°C (a measure of the volatile organic matter content of the solids).
1.2	Type of sample	Sewage and waterworks sludge and related solids.
1.3	Basis of the methods	 (a) Total solids content — gravimetrically by evaporation and drying in a tared dish. (b) Loss on ignition of dry residue — gravimetrically following ignition at 550°C. of the residue from the total solids determination
1.4	Range of application	(a) 0-100% W/W total solids.(b) 0-100% W/W loss on ignition as % of total solids.
1.5	Standard deviation	 For both determinations precision and accuracy are sample-type dependent. (a) Relative standard deviation usually better than 10% but increasing as solids content falls (see 1.7 below). (b) Relative standard deviation usually better than 20% (see 1.7 below).
1.6	Limit of detection	0.1% (based on the criterion of detection).
1.7	Bias	 (a) If the dry residue is hygroscopic there will be a tendency for results to be slightly high. See also Section 3 for an effect which may cause negative bias. (b) If used as a measure of organic matter the bias is sample-type dependent; many inorganic compounds including some metal oxides and salts volatilize or decompose below 550°C causing a positive bias.
1.8	Interferences	See Section 3.
1.9	Time required for analysis	The overall time for all determinations is sample- type dependent but under 7 h per batch. Batch size is dependent on the apparatus used but 20 and over could be achieved.

2. Principle

- (a) The total solids content of sludge is determined gravimetrically by drying at 105°C to constant weight.
- (b) The loss on ignition of dry residue is also determined gravimetrically. The residue in the dish from the total solids content determination is ignited in a furnace at 550°C to constant weight.

3. Interferences

- (a) The presence of substances which are volatile at temperatures below 105°C or which decompose at temperatures below 105°C to form volatile compounds will give falsely low results for the total solids determination.
 - (b) There are theoretically no interferences with the loss on ignition determination. However, for many purposes this determination is used as a measure of the organic solids content. In these circumstances the loss of volatile inorganic substances may occur giving high results and there may be incomplete combustion of all organic substances giving low results.
- 4. Reagents

No reagents are required.

5. Apparatus

- 5.1 Silica or platinum dishes or crucibles suitable for ignition at 550°C
- 5.2 Water bath (boiling)
- 5.3 Ovens controlled at $105 \pm 2^{\circ}$ C and
- 5.4 Muffle furnace controlled at 550 ± 20 °C
- 5.5 Analytical balance capable of weighing to \pm 0.1 mg
- 5.6 Desiccator

For notes on care of platinum ware see Ref.2.

6. Sampling

It is most important that the sample is truly representative of the sludge concerned (see publication in this series Ref 10). Sludge samples may change composition through microbiological activity and it is very important that they are analysed as soon as possible.

Step	Procedure	Notes
	Total Solids content	
7.1	Weigh a suitably sized dish. Let the weight be W ₁ g.	(a) Ensure that a representative subsample is taken. The weight of subsample required will vary with
7.2	Transfer between 10 and 50g of the sample into the tared dish (Note a) and reweigh. Let this weight be W_2g .	the anticipated total solids content of the sample. Sludge cake should be broken up.
7.3	Place the dish on a boiling water bath and allow the contents to evaporate to near dryness. (Note b)	(b) If a forced ventilation oven is available this step may be carried out in the oven.
7.4	Remove the dish from the water bath, blot any water from the bottom and put it into an oven at 105°C for at least 1 hour. Remove the dish, place in a desiccator until cool and reweigh.	(c) Retain the dish and solid for the determination of loss on ignition of dry residue.
7.5	Replace the dish in the oven for a further 15 minutes, cool and reweigh. Repeat this step until a constant weight is obtained. Let this weight be W ₃ g (Notes c and d).	(d) Some samples may continue to lose weight in which case constant weight may be difficult to obtain.

Calculation

7.6 Total solids content =

$$\frac{W_3 - W_1}{W_2 - W_1} \times 100\% \text{ W/W}$$

(as a percentage of the wet sludge)

Loss on ignition of dry residue

- 7.7 Place the dish and solid from Step 7.4 in the muffle furnace at 550° C (with the door open initially) for at least 30 min. (note e). Remove the dish, place in an oven at $120 \pm 20^{\circ}$ C, place in a desiccator until cool and reweigh (note f).
- the final weighing in Step 7.5 may be used.

(e) If short of sample and time permits the dish from

- 7.8 Replace the dish in the furnace for a further 10 minutes, cool and reweigh. Repeat this step until constant weight is obtained. Let this weight be W₄g.
- (f) Great care must be taken to avoid solids loss during the transfer operations of this step.

Calculation

7.9 Loss on ignition =

$$\frac{W_4 - W_3}{W_3 - W_1} \times 100\% \text{ W/W}$$

(as a percentage of the total solids content)

8. Sources of Error

The attention which it is necessary to pay to sources of error depends on the accuracy required of the analytical results. See Section 3 for the effect of interfering substances; further information is given below.

- 8.1 Some of the ammonium salts present will be lost during the determination of total solids. The remaining ammonium salts will be included with the loss on ignition.
- 8.2 All sewage sludges contain iron and most of the iron will be present as ferrous iron. During ignition ferrous iron will be oxidized to ferric iron with a consequent increase in weight and therefore a reduction in the loss of ignition.
- 8.3 If a sludge has been conditioned with lime the calcium hydroxide or calcium oxide present will be available to combine with any sulphur oxides liberated during ignition.
- 8.4 If adequate air is not present during ignition, elemental carbon may be left which could introduce reducing conditions. The oxidation states of some metals may be lowered and, eg, metal sulphates may be reduced to sulphides.

9. Checking the Accuracy of Analytical Results

Once the methods have been put into routine operation some factors may subsequently adversely affect the accuracy of analytical results. It is recommended that experimental tests of the accuracy should be made regularly. As a minimum, however, it is suggested that a typical sample be analysed in duplicate at the same time. The results should be plotted on a quality control chart which will enable reproducibility to be checked.

Determination of the Suspended Solids Content of Sewage and Waterworks Sludges and Related Solids

Method A Filtration - Direct Method B Centrifugation/Filtration - Indirect

Performance Characteristics of the Methods

1.1	Parameter determined	Suspended solids content.
1.2	Type of sample	Sewage and waterworks sludge suspensions and related solids.
1.3	Basis of the method	Method A should be used for any suspensions which filter easily and Method B should be used for any which do not filter easily. (a) Gravimetrically following filtration through a glass fibre or cellulose acetate filter medium and drying. (b) Gravimetrically by the difference between the total solids content of the suspension and the total solids content of the centrate/filtrate after centrifugation/filtration of the suspension.
1.4	Range of application*	(a) 500-10,000 mg/l. (b) Greater than 10,000 mg/l.
1.5	Standard deviation	 (a) Relative standard deviation is usually better than 5%. (b) If the sample is centrifuged to separate, the relative standard deviation is usually better than 20%, if the sample is filtered to separate, the Relative standard deviation is usually better than 15%.
1.6	Limit of detection	100 mg/l.
1.7	Bias	Sample dependent. Samples which are hygroscopic tend to give high results. See also section 3 for sample types tending to have negative bias.
1.8	Interferences	See Section 3.
1.9	Time required for analysis	The overall time for these methods is sample dependent, but should be under 7 h per batch. Batch size is dependent on equipment availability and capacity.

^{*} The ranges of application are given as a guide. Ease of filtration is a more important criterion.

2. Principle

- (a) The suspended solids content of "thin" sludges (those containing less than about 10,000 mg/l suspended solids) or other readily filtered suspensions is determined by filtering the suspension through a tared filter medium, drying in an oven at 105°C and reweighing to constant weight.
- (b) The suspended solids content of sludges and other suspensions which do not filter readily is determined by an indirect method. The total solids content of the sludge is

determined by drying a sample in the oven at 105°C to constant weight. A further portion of the sludge is centrifuged/filtered and the total solids of the centrate/filtrate determined by drying in an oven at 105°C to constant weight. The suspended solids concentration = (concentration of total solids of sludge) — (concentration of solids of centrate/filtrate).

3. Interferences

The presence of substances which are volatile at temperatures below 105°C or which decompose at temperatures below 105°C to form volatile compounds will give falsely low results.

4. Reagents

Distilled or deionized water is required. No other reagents are needed.

5. Apparatus

5.1 General

Analytical balance capable of weighing to \pm 0.1 mg Oven controlled at 105 \pm 2°C Desiccator

5.2 A. Filtration Method

Glass fibre or cellulose acetate filter medium
3 component Hartley-type filter funnels
Buchner filter flasks
70 mm diameter is suitable

5.3 B. Centrifugation/Filtration Method

Centrifuge capable of at least 1,500 r.c.f. with tubes of at least 100 ml capacity or filtration apparatus

Silica or platinum dishes

Water bath (boiling)

6. Sampling

It is most important that the sample is truly representative of the sludge concerned (see publication in this series Ref 10). Sludge samples may change composition through microbiological activity and it is important that they are analysed as soon as possible.

to retain sample, then it is either placed on the

special rack or between filter papers, which are

carefully removed later.

7. Analytical Procedure A.

Filtration Method

Step	Procedure	Notes
7.1	Wash the filter medium by the following procedure: Insert a glass fibre or cellulose acetate filter medium into the funnel assembly, and using slight suction wash with approximately 100 ml of water. Carefully remove the filter medium, (note a) and dry in the oven for approximately 1h at 105°C. Cool the filter medium in a desiccator and weigh. Let the weight be W ₁ g.	(a) Care must be taken in placing the filter media in the oven. They may be placed on a specially designed rack or placed between filter papers.
7.2 Replace the filter medium in the funnel assembly and moisten with a little water. Measure a suitable volume, V ml, (up to 100 ml) of the well mixed sample into a measuring cylinder (note b). Using slight suction filter the sub-sample ensuring that all the solids are transferred to the filter medium. Wash the residue three times with approximately 10 ml of water, allowing it to drain free of water after each wash.		(b) Ensure that a representative sub-sample is taken. The volume of sub-sample required will vary with the anticipated suspended solids content but should be such that at least 1 mg of solid is weighed following drying, but preferably 1g.
7.3	Remove the filter medium, (note c) and dry in the oven at 105°C for at least 1 hr. Cool in a desiccator and	(c) Care must be taken in placing the used filter medium in the oven. First the medium is folded

weigh.

Step	Procedure	Notes
7.4	Replace the filter medium in the oven at 105°C for a further 15 minutes. Cool in a desiccator and reweigh. Repeat this step until a constant weight is obtained (note d). Let this weight be W ₂ g.	(d) Some samples may continue to lose weight and a constant weight may be difficult to obtain.
	Calculation	
7.5	Suspended solids content	
	$= \frac{W_2 - W_1}{V} \times 10^6 \text{ mg/l}$	
8. <i>A</i>	Analytical Procedure B.	
Cent	trifugation/Filtration Method	
Step	Procedure	Notes
	Total Solids content of suspensions	
8.1	Weigh a suitably sized dish. Let the weight be W_1g .	(a) Ensure that a representative sub-sample is taken.
8.2	Transfer between 10 and 50 g of the sample into the tared dish (note a) and reweigh. Let this weight be W_2g .	The weight of sub-sample required will vary with the anticipated total solids content of the sample.
8.3	Place the dish on a boiling water-bath and allow the contents to evaporate to near dryness (note b).	(b) If a forced ventilation oven is available this step may be carried out in the oven.
8.4	Place the dish for at least 1 h in an oven at 105°C. Remove the dish, place in a desiccator until cool and reweigh.	(c) Some samples may continue to lose weight and a constant weight may be difficult to obtain.
8.5	Replace the dish in the oven for a further 15 minutes, cool and reweigh. Repeat this step until a constant weight (note c) is obtained. Let this weight be W ₃ g (note d).	(d) If required, retain the dish and solid for the determination of loss on ignition of dry residue.
	Calculation	
8.6	Total solids content of suspension =	
	$\frac{W_3 - W_1}{W_2 - W_1} \times 100\% \text{ W/W}$	
	Total Solids content of the centrate/filtrat	te.
8.7	Centrifuge or filter sufficient sludge to give at least 50 ml of clear centrate/filtrate (see section 9).	
8.8	Weigh a suitably sized dish. Let the weight be W ₄ g.	
8.9	Transfer approximately 50 ml of the clear centrate/filtrate (note e) to the dish and weigh. Let the weight be W_5g .	(e) For centrates take care to exclude floating matter
8.10	Repeat steps 8.3 to 8.5 inclusive. Let the constant weigh be W_6g . (note d).	

Calculation

8.11 Total solids content of centrate/filtrate.

$$\frac{W_6 - W_4}{W_5 - W_4} \times 100\% \text{ W/W}$$

Calculation of suspended solids content

8.12 Suspended solids content, per cent w/w = (total solids content of suspension — total solids content of centrate/filtrate).

9. Sources of Error

The attention which it is necessary to pay to sources of error depends on the accuracy required of the analytical results. See Section 3 for the effect of interfering substances.

Method B assumes that the total solids content of the centrate/filtrate is equal to dissolved solids content of the original suspension. It may not always be possible to obtain a clear centrate or filtrate because of the presence of colloidal matter.

10. Checking the Accuracy of Analytical Results

Once the method has been put into routine operation many factors may subsequently adversely affect the accuracy of analytical results. It is recommended that experimental tests of the accuracy should be regularly made. As a minimum, however, it is suggested that a typical sample be analysed in duplicate at the same time. The results should be plotted on a quality control chart which will enable reproducibility to be checked.

Determination of the Filterability of a Sludge Using the Capillary Suction Time (CST) Method

I. Performance Characteristics of the Method

1.1	Parameters determined	Capillary suction time (CST).	
1.2	Type of sample	Any suspension of solid particles in an aqueous liquid, excepting very thin sludges where the concentration of suspended solids is less than 0.5 per cent or those sludges which have a CST very close to that of water alone.	
1.3	Basis of the method	Measurement of the time for a given area of paper to become wetted with filtrate from a sludge using a standard apparatus.	
1.4	Range of application	CST depends on the size of reservoir, and on the nature of sludge and paper. 10-mm diam. reservoir 9-3000s 18-mm diam. reservoir 5-3000s The lower figure in each case is the CST of water.	
1.5	Standard deviation	Relative standard deviation usually better than 10 per cent.	
1.6	Interferences	See Section 3.	
1.7	Time required for analysis	Overall time is dependent on the time to determine the concentration of solids in the sample (see method in this series). Up to 5 minutes for individual CST measurements.	

2. Principle

Filtrate is withdrawn from a sludge sample placed in contact with an absorbent filter paper by the capillary suction exerted by the paper. The rate at which filtrate wets the filter paper is mainly dependent upon the sludge filtrability. The time to wet a standard area of paper between two sets of electrodes which start and stop an electrical clock is the Capillary Suction Time (CST). This time is also dependent on the size of the sludge reservoir used and on the concentration of suspended solids of the sludge, which latter therefore has to be determined.

Interferences

Some dissolved chemicals in the filtrate of the sludge, particularly unadsorbed polyelectrolyte, may adsorb on to the CST paper and produce a falsely high CST.

4. Reagents

Sodium Chloride, reagent grade, may be needed.

5. Apparatus

- 5.1 The standard Capillary Suction Time apparatus (Ref. 3) (Fig 1) with 2 different sized sludge reservoirs of 18 mm and 10 mm diameter. (It is recommended that the counting mechanism be checked at intervals against a stop watch and if necessary adjusted to give deviations not greater than $\pm 1\%$).
- 5.2 A supply of standard CST papers. (Whatman No. 17 Chromatography Grade Filter paper of nominal thickness 0.88 mm and weight per unit area of 440 g/m²).

- 5.3 Two 250 ml low form eg. ISO draft spec. 522 glass beakers.
- 5.4 Paper tissues.
- 5.5 A 100 ml wide-mouthed sample bottle.
- 5.6 A fluorescent light box (not essential).
- 5.7 Desiccator, to store the CST papers.

6. Sampling

It is important that the sample is truly representative of the sludge concerned (see publication in this series Ref 10). Sludge samples may change composition through microbiological activity or temperature variation. It is important therefore that the samples are analysed as soon as possible after collection. It is necessary to know the time elapsed from the time of taking the sample and the change in temperature so that the effect of storage on the filtrability can be assessed. Coarse solids should be removed by passing the sample through a 4-mm mesh sieve before testing.

7. Procedure

Step	Procedure	Notes
7.1	Prepare the CST apparatus (after ensuring that the electrical probes are clean) with the probes resting on the rougher side of the CST paper. Initially use the 18 mm diameter sludge reservoir (note a).	(a) The polymethylmethacrylate (or similar) block carrying the paper and reservoir may be placed on a light box to aid observation of the water front.
7.2	Pour tap water into the sludge reservoir, observe advancing water interface and check that the clock starts and stops satisfactorily. If the conductivity of the tap water is too low to trigger the timing mechanism, dissolve sufficient sodium chloride in the water to raise the conductivity high enough to work the mechanism (over about 500 µmhos/cm). Record the CST of water alone (note b).	(b) This is for reference. The CST of a sludge sample is not very meaningful if it is close to that of water.
7.3	Remove and discard wetted CST paper. Clean and dry reservoir and acrylate block with tissues. Reassemble apparatus.	
7.4	Thoroughly mix sample by carefully pouring sludge sample from one beaker to another (notes c and d) and immediately pour a sub-sample into the sludge reservoir.	 (c) The filtrate from the sludge must have a conductivity sufficiently high (ca 500 μmhos/cm) to trigger the electrical clock. The conductivity of sewage sludges is usually high enough, but some waterworks sludges may require addition of a small amount of ionizable material, such as sodium chloride, to raise the conductivity. (d) Sludge properties may be either temporarily or permanently altered if shaken.
7.5	Record CST. Let this be t ₁ seconds.	
7.6	Repeat steps 7.3–7.5. Let second CST be t_2 seconds (note e).	(e) If t ₁ and t ₂ differ by more than 20 per cent, the clock may not have operated correctly, or it may not be possible to get a representative subsample of sludge. Stir sample and repeat procedure.
7.7	After transferring the remaining sludge sample from one beaker to another pour a portion of the sludge into the 100-ml sample bottle for subsequent determination of suspended-solids content as described in this booklet.	

7.8	If the CST determined using the 18-mm diameter sludge reservoir has a value of less than 100 s, it is desirable to repeat the procedure using the 10-mm	

7.9 Quote results as CST $\frac{t_1 + t_2}{2}$ seconds

at a suspended-solids content of C per cent dry matter (stating the size of reservoir). (note b).

8. Sources of Error

Procedure

reservoir.

Step

The attention which it is necessary to pay to sources of error depends on the accuracy required of the analytical results. See Section 3 for the effect of interfering substances; further information is given below.

Notes

- 8.1 Because the CST of a sludge with less than 0.5 per cent dry solids is not very meaningful, (see Section 9 below), it is usually desirable to remove as much water as possible from thin sludges by settlement before the determination. Ideally the supernatant liquor removed should not contain any suspended solids, but in practice the removal of some solids will tend to reduce the CST.
- 8.2 Biological sludges are often very unstable and several determinations of filtrability during a storage period may be desirable to assess the effect of time. Samples for analysis should be taken just prior to analysis.
- 8.3 If left for some time on the light box, the temperature of the block will rise, which will affect the CST.
- 8.4 Shear, in for example stirring, can affect CST and it is desirable to use a standard stirrer (see method in this series for determining coagulant demand).
- 8.5 CST is dependent on the solids content of the sample, but the results may be expressed in terms of the more absolute parameter, specific resistance to filtration after constructing a calibration graph of the CST against the product of suspended solids of the sludge and the specific resistance to filtration (see method in this series).
- 8.6 Batches of CST paper may vary slightly in absorptive power and thickness. Therefore it is desirable to note the batch number of paper being used, and where specific resistance to filtration is required to recalibrate for each batch of paper.

9. Interpretation

CST are useful for controlling the dosing of sludges prior to filtration. CST below a critical value, dependent on conditioner and filter, invariably give good quality filter cake.

10. Accuracy of the CST Determination

The principal factors affecting the CST of a sludge will be the filtrability (the specific resistance to filtration) and the suspended solids of the sludge. Changes in either of these two parameters can cause the CST to alter by a factor of 100 or more, with sludge varying in CST from 5 seconds to over 3000 seconds.

There are a number of other factors which can also affect the CST, fortunately, usually to a much lesser degree than the filtrability or suspended solids of the sludge. They may be either associated with the instrument together with the conditions surrounding its use, or pertaining to the sludge being used.

Variation in CST results due to the instrument and operating conditions may be caused by:

A variation in the thickness and weight per unit area of the CST paper. (The paper used in the CST apparatus is actually a Whatman No 17 chromatography grade filter paper of nominal thickness 0.88 mm and weight per unit area of 440 g/m^2 .)

A variation in the temperature of the sludge and the humidity of the atmosphere in which the CST papers are stored.

A variation in the accuracy of the timing mechanism in the CST counter, or the physical dimensions of the CST perspex block.

The variability associated with the sludge is dependant on:

The degree to which the sludge sample is homogenous, (such that a representative sample can be poured into the CST reservoir).

The extent to which the sludge settles during the time over which the CST measurement takes place.

The viscosity of the filtrate being withdrawn from the sludge.

If necessary some of the variability associated with the CST apparatus, or how it is used, can be overcome by measuring the CST at a fairly constant temperature and keeping the CST papers in an atmosphere of fairly constant humidity, such as pertains in a centrally heated laboratory, or a desiccator. However, variation in CST due to either temperature or humidity is fairly small with a reduction in temperature of 16°C causing a 20% increase in CST, while a reduction in relative humidity by 80% at constant temperature causes a similar increase.

The variation in CST due to variation in paper thickness and weight per unit area is best overcome by using CST papers from the same manufacturing batch for one whole series of tests. (Each production batch of papers has its own number indicated on the outside of the box of CST papers). This is desirable because the variation within each batch of paper manufactured at any one time is fairly small, (with coefficients of variation in thickness being of the order of 2 per cent or less) while the variation between different batches can be considerable. If this is the case when the thickness of two batches of paper vary by X per cent, then the difference in CST of a sludge between that given by each batch of paper has been found to be of the order of 2X per cent.

The counting mechanism of the CST apparatus should, of course, count in seconds. It is advisable to check occasionally the accuracy of this mechanism against a stop-watch. Should the counter be inaccurate by more than 1 per cent resetting of the internal variable potentiometer should restore the accuracy of timing.

Inaccuracies due to the polymethylmethacrylate CST block are usually associated with probe corrosion. This may usually be overcome by gentle rubbing on a fine grade emery paper. Occasionally probes become damaged or deformed; should this occur replacement of the whole block should be made, because any shift in the position of the start and stop probes relative to one another, or to the sludge reservoir will substantially effect the results, as the CST is proportional to the fourth power of the radius of the circle in which the probe is set.

Variable and atypical CST results due to the sludge are usually due to lack of homogeneity of the sample, particularly after a conditioning chemical such as a polyelectrolyte has been added. Satisfactory sludge sampling is usually easier if immediately prior to pouring the sludge sample into the CST reservoir the sample is transferred between vessels. If liquor only is poured into the reservoir it is not valid to treat this result as the CST of the sludge, but the result can still be of considerable value. If the results of this CST, with liquor only in the reservoir, is different to that of water, it is usually indicative of surplus unadsorbed chemical, usually a polyelectrolyte being present in the liquor. The CST is then higher than that of water because of the increased viscosity of the liquor as compared with water, and the ability of the CST paper to adsorb the surplus polyelectrolyte.

The variation in CST will therefore be dependent upon the interaction of many different factors. The results for 2 different circumstances were as follows:

1) Anaerobic digested sludge (elutriated)

All the data is from one instrument being the average of 8 results in 6 series of tests. 3 series with rough side of paper uppermost, 3 series with smooth side uppermost.

Paper surface	smooth	smooth	rough	smooth	rough	rough
Individual means(S) Individual standard	201.1	202.5	201.4	204.6	193.3	205.5
deviation (S) Individual variance	11.86	12.38	14.90	15.98	21.11	43.04
(per cent)	5.87	6.11	7.40	7.81	10.9	21.0

2) Water only

Same instrument as 1, and similar series of tests

Paper surface	rough	rough	smooth	smooth	smooth	rough
Individual means(S) Individual standard	8.04	7.85	8.23	7.90	8.29	8.01
deviation (S) Individual variance	0.3419	0.5099	0.5365	0.5292	0.6129	0.6977
(per cent)	4.25	6.50	6.52	6.70	7.39	8.71

Conclusion

The typical range of variances in CST results is from 6 to 10 per cent but other factors can cause a considerable increase in this figure.

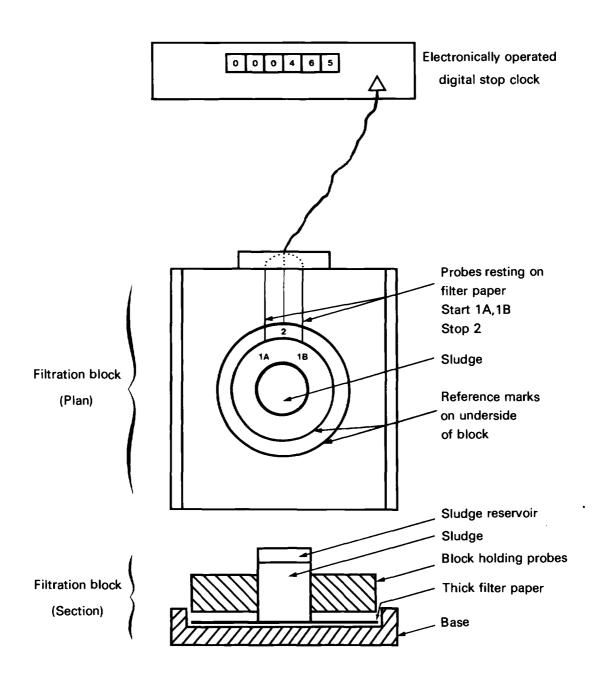


FIGURE 1 DIAGRAM OF THE CST APPARATUS

Determination of the Conditionability of a Sewage and Waterworks Sludge

1. Introduction

It is usually necessary to condition, that is improve the filtration characteristics of, a sewage or waterworks sludge before mechanical dewatering. Conditioning may also be advantageous before dewatering on drying beds. Conditioning may be effected by physical means (heat treatment, freeze/thaw) or by the addition of chemical coagulants. The object of this method is to facilitate the choice of the type and dose of chemical for any particular sludge and any particular type of mechanical dewatering equipment. The method is described as the determination of a characteristic of the sludge, that is its amenability to conditioning, or the quantity of chemical required to achieve a given filtrability measured either by specific resistance to filtration (SRF) or Capillary Suction Time (CST). However, the filtrability required will depend on the mechanical dewatering equipment to be used. The type of equipment will also define characteristics of the conditioned sludge other than filtrability, such as the strength of the floc and its susceptibility to shear in pumps, pipework, etc. Approximate values of specific resistance to filtration and other parameters for satisfactory performance of various types of mechanical dewatering equipment are given in Table 1.

2. Performance Characteristics of the Method

2.1	Parameters determined	The amount and type of chemical required to improve the filtrability of a sludge to a value such that the sludge is suitable for mechanical dewatering.
2.2	Type of sample	Sewage, waterworks or industrial sludge.
2.3	Basis of method	The filtrability of a sludge is measured using the Capillary Suction Time (CST) apparatus before and after varying quantities of different coagulating chemicals have been added to the sludge.
2.4	Time required for analysis	2-4 hours per chemical.

3. Principle

Samples are conditioned with various quantities of different chemicals and compared with a control to which the same quantity of water has been added. Filtrability is assessed by determining the CST of the conditioned sludge. Floc strength is assessed by determining the CST after different periods of shear in a standard laboratory stirrer. The sludge is said to be conditioned when the filtrability and floc strength are such that the sludge can readily be dewatered using the chosen apparatus. The amount of chemical required to achieve this conditioning is termed the coagulant demand. Either a single chemical or a combination of chemicals may be used.

4. Reproducibility

Tests by Southern and other WAs indicate that with the exception of very wet sludges or those high in solids that float, this test can indicate optimum plant conditions within 20%.

5. Reagents

The single chemical systems which can be used for conditioning sewage sludge are:

Aluminium chloride Aluminium chlorohydrate Ferric chloride Cationic polyelectroytes When dual chemical systems are used lime (either calcium hydroxide or calcium oxide) is almost invariably one of the chemicals. The second chemical can be any of the above or more usually a ferrous salt such as ferrous sulphate.

For conditioning waterworks sludges, anionic polyelectrolytes may be more suitable.

6. Apparatus

- 6.1 The standard Capillary Suction Time apparatus see previous method in this booklet with 2 different size sludge reservoirs of 18-mm and 10-mm diameter. (Multiple units will speed up the procedure.)
- 6.2 A supply of standard CST papers. (Whatman No 17 Chromatography Grade filter paper of nominal thickness 0.88 mm and weight per unit area of 440 g/m².)
- 6.3 A standard laboratory stirrer (designed by WRC) (Ref. 3)
- 6.4 Six 250-ml low form glass beakers, calibrated at 50-ml intervals
- 6.5 Two 2-1 glass beakers
- 6.6 A 20-ml calibrated pipette
- 6.7 A 20-ml graduated pipette
- 6.8 Rubber-bulb pipette filler
- 6.9 Paper tissues
- 6.10 Wide-mouthed 100-ml sample bottles
- 6.11 Fluorescent light box (not essential)
- 6.12 Desiccator to store CST papers

7. Sampling

It is most important that the sample is truly representative of the sludge concerned (see publication in this series Ref 10). Samples may change composition through chemical or microbial activity and it is important that the test is carried out as soon after sampling as possible.

8. Procedure

Step	Procedure	Notes	
8.1	Prepare at least 100 ml of a standard solution of the chemical to be tested (note a.)	(a) The strengths of the standard solur as weight per cent volume are:	tions expressed
		Aluminium chlorohydrate	1% Al ₂ O ₃
		Aluminium chloride	1% A1 ₂ O ₃
		Ferric chloride	5% FeC13
		Ferrous sulphate	10% FeSO ₄
		(Prepared with deoxygenated	
		distilled water acidified to pH 1)	
		Polyelectrolyte (solid)	1.0% product
		Polyelectrolyte (liquid)	1.0% product

Step	Procedure	Notes
8.3	Pour the sludge into another 2-1 beaker and back again. Repeat until the sludge is thoroughly mixed, then pour ca. 50 ml into a wide-mouthed 100-ml sample bottle for determination of suspended solids (see method in this booklet), and measure at least five 100-ml portions of sludge directly into calibrated 250-ml beakers.	
8.4	Place one 250-ml beaker containing 100 ml of sludge under the standard laboratory stirrer.	
8.5	Fill a 20-ml pipette with distilled water, press the start button to commence the 10-s stirring period and immediately start to empty the pipette into the sample of sludge (note b).	(b) The pipette should always drain sufficiently rapidly to empty by the end of the 20-s stirring period. A standard pipette can be modified by removing part of the tip.
8.6	Lower the beaker so that the stirrer blade is just above the surface of the sludge, and switch on stirrer for a brief period, to remove adhering sludge. Remove beaker from stirrer and pour sludge into a clean 250-ml beaker and back again, and immediately pour a sample into the CST reservoir (note c).	(c) For most sludges the 18-mm CST reservoir is used. For details of the use of the CST apparatus, see method in this booklet.
8.7	Replace remaining sludge under stirrer and stir for a 20-s period, then repeat Step 8.6.	
8.8	Repeat Section 8.7 for a further 30-s stirring period.	
8.9	Repeat Section 8.7 for a further 60-s stirring period (note d).	(d) The CST results for Sections 8.6 to 8.9 are for zero dose of chemical for 4 different stirring periods, and are control data.
8.10	8.10 Estimate (or determine by direct measurement) the concentration of suspended solids in the sludge. Let the value be y per cent dry matter. Measure 8y ml of the chemical solution prepared in Section 8.1, using a calibrated pipette, into an empty 250-ml beaker. Dilute to a total volume of 40 ml by adding (40-8y) ml of distilled water (note e). (e) If a two-chemical system is to be one of the chemicals is lime, pretruence unconditioned sludge with lime, cent suspension, to give a typical per cent weight of lime (calcium weight of dry sludge solids.	
8.11	Using half of the 40-ml portion of diluted solution prepared in Section 8.10 fill a 20-ml pipette, and set aside the remaining 20 ml of solution for subsequent	(f) This dose is the maximum to be used (usually expressed as weight per weight of dry sludge solids). For each chemical this will be:
	use (note f).	Aluminium chloride and chlorohydrate 4% Al ₂ O ₃ Ferric chloride 20% FeCl ₃ Polyelectrolyte 0.4%
8.12	Repeat Steps 8.4 to 8.9, using the 20 ml of diluted solution from 8.11 instead of distilled water (note g).	(g) If a second coagulant is to be used add after removing beaker from stirrer in Step 8.5. Mix by pouring from beaker to beaker.
8.13	To the 20 ml of diluted solution left over from Section 8.11 add 20 ml of distilled water to produce 40 ml of solution of half the strength of that used in Section 8.12.	
8.14	Repeat Sections 8.4 to 8.9 using the diluted solution prepared in Section 8.13 (note h)	(h) This dose represents half of the maximum dose.

prepared in Section 8.13 (note h).

8.15

Section 8.14.

To the 20 ml of dilution solution left over from

Section 8.13 add 20 ml of distilled water to produce 40 ml of solution of half the strength of that used in

prepared in Section 8.15 (note i). For completeness this progressive dilution	Step	Procedure	Notes
conditioned sludge approaches the CST of the	8.16	•	technique can be continued until the CST of the conditioned sludge approaches the CST of the unconditioned sludge, as measured in Steps 8.5 to

9. Interpretation of Results

Results may be plotted as CST against period of stirring in the standard apparatus. Some examples are given in Fig. 2. For a well-conditioned sludge with a high floc strength the line should be low and almost horizontal (Curve A). For a filter belt press the initial CST of the conditioned sludge should approach that of water, and the curve should rise steeply (Curve B).

If the curve falls initially (Curve C), it indicates inadequate mixing of the sludge and conditioner, and may also indicate that the concentration of polyelectrolyte is too high.

For assessment of the optimum concentration of any given chemical, concentration should be plotted against CST measured after a period of stirring appropriate to the type of dewatering being considered. For example, sludge fed to a drying bed, rotary vacuum filter or filter belt press undergoes little shear, and a period of stirring of 0-10 s is appropriate. Sludge pumped under pressure into a filter press will experience shear equivalent to 20-40 s in the laboratory stirrer, depending on the type of pump, length of pipework, etc. Worn pumps, excessive lengths of pipe or numbers of valves and bends may exert a level of shear equivalent to much greater than 40 s in the laboratory stirrer.

A typical graph of concentration of chemical against CST after stirring for a period of 40 s is given in Fig 3. In order to compare different chemicals, the cost of each chemical at its optimum dose can be compared, or the cost of each chemical can be plotted against CST on one graph (Fig. 4).

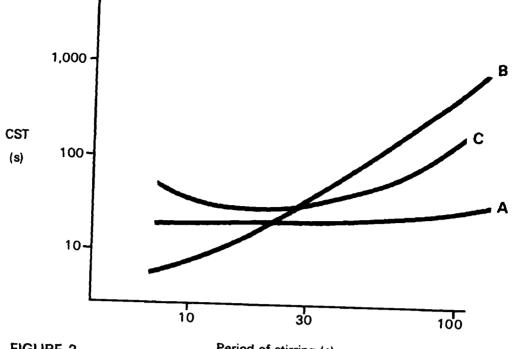
10. Sources of Error

- 10.1 Conditioning may be strongly dependent on pH value, alkalinity, presence of dissolved salts, and concentration of very fine particles, all of which may change owing to microbial action.
- 10.2 It is very difficult to maintain homogeneity in a superflocculated sludge, and the initial CST may not be representative of the whole sample.
- 10.3 Some chemical solutions will lose their effectiveness with time, especially if made up with tap water. It is recommended that new solutions are made up regularly. It is also recommended that polyelectrolyte solutions at a concentration of 1% are prepared using de-ionized water. They can be stored for up to week in the dark and diluted to 0.1% for the test.
- 10.4 There are a number of sources of error associated with the use of the CST apparatus, and reference should be made to the method in this booklet.

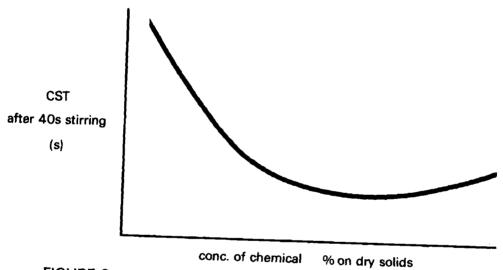
Table 1. Approximate values of specific resistance to filtration and other parameters for satisfactory performance of various types of mechanical dewatering equipment

Properties required of conditioned sludge to ensure satisfactory dewatering						
Method of dewatering	Specific resistance to filtration (10 ¹² m/kg)	CST at 5% solids (18-mm reservoir) (s)	Floc strength			
Drying beds with underdrains	30	300	Not very important			
Filter press	1	15	Strong			
Vacuum filter	4	40	Medium. Flocs should be strong enough to avoid cake drop-off			
Filter belt press*	-	10	Weak			
Centrifuge*	4	40	Strong			

^{*} The method does not give all the information required for the selection of chemicals for filter belt pressing or centrifugation. In the former case super-flocculation is required and tests to determine the rate of drainage and the final solids content of the cake should be carried out. For centrifugation pilot-scale tests are normally preferred.



Period of stirring (s) FIGURE 2



conc. of chemical FIGURE 3

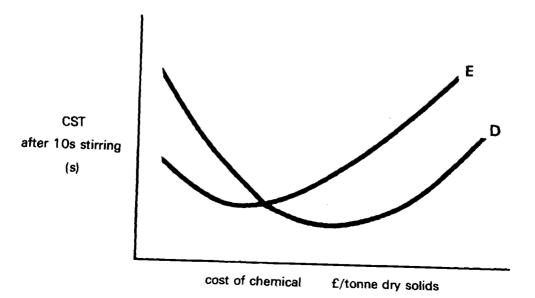


FIGURE 4

Methods for the Initial Selection of Chemicals for Use as Settlement Aids in the Treatment of Wastewater

1. Introduction

Chemicals are being employed increasingly in wastewater treatment to enhance settlement by coagulation and flocculation.

"Coagulation" and "flocculation" are terms which are often used loosely and interchangeably. Coagulation is defined as the neutralisation of surface charges on the particles to be removed by settlement. The small non-settleable particles present in sewage carry negative charges causing repulsion between particles. Once coagulation has been effected flocculation can occur. This involves the agglomeration of the coagulated colloids or of larger particles by polymer bridging to form larger, more rapidly settling flocs. Flocculation is brought about by gentle agitation which is sufficient to cause the particles to collide but not violent enough to cause floc break up.

The methods are designed to screen available chemicals for their applicability as settlement aids for a given wastewater before carrying out full-scale tests. Laboratory-scale settlement tests using vessels of two different sizes are described. As it is difficult to simulate full-scale mixing conditions in the laboratory, it is not expected that the tests will enable the removal of suspended solids on the full scale to be predicted accurately. However, examination of the relative effect of different chemicals should enable those chemicals to be selected which warrant trial on a full scale.

Similarly, it is important to be able to examine the relative characteristics of sludges produced by different chemicals, as the use of a settlement aid may increase the quantity of sludge produced, and the cost of sludge treatment and disposal may be a significant proportion of the total cost of wastewater treatment. Thus the size of vessel in one method is such that a usable quantity of sludge is produced.

2. Performance Characteristics of the Methods

2.1	Parameter determined	Quantity of chemical for a given enhancement of settlement.
2.2	Type of samples	Sewages or industrial wastewaters.
2.3	Nature of method	Qualitative.
2.4	Basis of method	Comparisons of effects of different concentrations of chemicals on: A. Quality of supernatant liquor in 1-litre beakers. B. Quantity and quality of sludges produced in 25-litre vessels.
2.5	Time required for	 A. 3 hours excluding determination of suspended solids and other desired parameters. B. 4 hours excluding determination of suspended solids and other desired parameters.

3. Principle

Samples are treated with different quantities of chemical and compared with a control containing no chemical, under standardized conditions in a jar test. After flocculation the samples are allowed to settle and the effectiveness of the chemical is judged by the qualities of the supernatant liquor and sludge. The principle is one of comparison on a small scale to give an approximate ranking before testing on a full scale. The results are not intended to predict a particular removal of suspended solids for a given dose of chemical.

4. Reagents

- 4.1 Settlement aids. These will usually include quick or hydrated lime, iron and aluminium salts and ranges of anionic and cationic polyelectrolytes.
 - 4.2 Acetone for initial wetting of solid polymer products.
 - 4.3 Acid and alkali for adjustment of pH value.
 - 4.4 De-ionized water.

Solutions to be made up according to the manufacturers' recommendations.

Method A

5. Apparatus

5.1 The principal item is a laboratory apparatus incorporating a bank of stirrers driven at a number of set speeds. Each stirrer consists of a vertically mounted shaft with a two-bladed paddle. Each blade is 35-mm wide and 50-mm deep (see Fig. 5). It is recommended that the apparatus has positions for six samples to be stirred simultaneously at the same speed. The stirrer speeds should be reproducible to within $\pm 5\%$ and all the speed settings needed by the test clearly marked on the speed controller. The stirrers should be located centrally in the beakers with the bottom edge of the paddle being 20 mm from the bottom of the beaker. Illumination of the samples can be helpful (see Section 9.2.)

Other equipment needed:

- 5.2 1-litre beakers "tall-form"
- 5.3 pH meter
- 5.4 Thermometer
- 5.5 Stop watch
- 5.6 1-litre measuring cylinders
- 5.7 Pipettes or syringes
- 5.8 Dosing rack
- 5.9 Apparatus for determination of suspended solids (see method in this booklet), or turbidity, or other desired parameter.

Method B

5.10 The essential equipment consists of two 25-litre test vessels* mounted in a suitable framework which holds the vessels and provides a support for mounting a variable-speed stirrer over each of the vessels. Each test vessel should be marked to indicate the level equivalent to 20 litres. Provision should be made to remove both supernatant liquor and settled sludge. This is conveniently done by inserting two vertical drain tubes of differing lengths (see Fig. 6) through the apex of the conical section. The longer pipe projects to a point above the sludge level and the other is fitted flush with the apex of the conical section. The upper pipe is used to drain supernatant liquor after the settlement period and the lower pipe is used to take the settled sludge which has collected in the conical section.

- (i) fabricated as shown in Fig. 6.
- (ii) made from conically-topped polyethylene bottles,
- or (iii) borrowed from WRC, Stevenage Laboratory.

^{*} Note: The apparatus may be either:

Other equipment needed:

- 5.11 pH meter
- 5.12 Thermometer
- 5.13 Stop watch
- 5.14 Measuring cylinders
- 5.15 Pipettes
- 5.16 Apparatus for assessing thickenability (see method in this booklet)
- 5.17 Apparatus for determination of Capillary Section Time (CST) (Ref 3), Specific Resistance to Filtration (SRF) (Ref 7), or other desired parameters.

6. Sampling

It is most important that the sample is truly representative of the wastewater concerned (see publication in this series Ref 10). Samples may change composition through chemical or microbial activity and it is important that the test is carried out as soon after sampling as possible. If it is not possible to carry out the test immediately, samples should be stored at 4°C.

7. Procedure for Method A

Step	Procedure	Notes
7.1	Measure the temperature and pH value of the sample. See Section 9.2. Adjust pH value if required (note a).	(a) The pH value for optimum settlement may not be the same as that of the sample. It is normally not economic to adjust the pH value of sewage on the full scale.
7.2	Thoroughly mix the sample and pour 800 ml portions into the 1-litre tall-form beakers.	(b) The range of concentrations to be added may be recommended by the supplier of the chemical, or may be chosen after reference to the literature.
7.3	Place the beakers in the apparatus. Lower the paddles and start stirring at the appropriate flash-mixing speed shown below in Table 1. After 15 s add the chosen chemical to produce a range of concentrations in the sample (note b).	Lime is normally added in the range 100-1000 mg Ca(OH) ₂ /l sewage, iron and aluminium salts in the range 10-500 mg product/litre sewage and polyelectrolytes in the range 0.1-3 mg product/litre sewage. The procedure may have to be repeated using a narrower range of concentrations to enable the optimum concentration to be located more accurately.

Table 1.

Step		Hydrated lime	Aluminium and iron salts	Polyelectrolytes	
Flash mixing (Coagulation)	Speed Period	100 rpm 2 min	100 rpm 2 min	300 rpm ≠ 30 s	-
Flocculation	Speed Period	20 rpm 15 min	20 rpm 15 min	40 rpm 4 min	

[≠] Enhanced performance may be obtained at higher stirring speeds.

Step	Procedure	Notes
7.4	Stir the samples for the appropriate flash mixing period shown in Table 1.	
7.5	Reduce the speed to that appropriate for flocculation (shown in Table 1) and continue for the period shown (note c).	(c) If a second chemical, or coagulant aid is to be used, stir slowly for only 2 min (longer periods may be necessary for very dilute suspensions). Add the second chemical, stir at high speed for 30 s, and proceed with step 7.5.
7.6	Switch off and remove stirrers. Note approximate floc size (See Section 10 below). Remove beakers and allow samples to settle noting the settlement rate. Withdraw a 50- or 100-ml sample from the middle of the supernatant liquor (note d) for determination of suspended solids or other desired parameter (note e) after 60 min (note f).	 (d) Size of sample appropriate to SS and other desired parameters. (e) Visual examination of samples may allow rough screening at this stage. (f) When screening for rapid removal of solids, samples may be examined after 15 or 30 min.
7.7	Measure temperature and pH value again (See 9.2).	
7.8	Repeat the procedure 7.1 to 7.7 for the next chemical of interest (note g).	(g) If different samples of the wastewater have been used in the determination of the optimum concentration for a number of chemicals, the test should be repeated by adding this concentration of each chemical in step 7.3. to aliquot portions of a single sample of the wastewater.
7.9	To compare chemicals, rank them in terms of their effectiveness in enhancing the removal of suspended solids (compared with the control) at the optimum concentration (note h).	(h) There is much confusion in the expression of concentrations of chemicals. For the purpose of comparison the concentrations should wherever possible be expressed as Al for aluminium salts, For for iron salts, and Ca(OH) ₂ for hydrated lime. Where the product has an indeterminate chemical composition, it is recommended that concentration be expressed as weight of a specified active ingredient per unit volume of liquid. For polyelectrolytes, and products which are a mixture of salts, concentrations should be expressed as weight of product per unit volume of liquid. For operational purposes it may be more convenient to express all concentrations as volume or weight of product supplied per unit volume of wastewater.
8. F	Procedure for Method B	
Step	Procedure	Notes
8.1	Measure the temperature and pH value of the sample. Adjust pH value if necessary (note a).	(a) The pH value for optimum settlement may not be the same as that of the sample. It is normally not economic to adjust the pH value of sewage on the full scale.
8.2	Thoroughly mix the sample, and pour 20-litre	

(b) The optimum speeds for flash mixing and

being tested.

flocculation may vary over a range dependent upon the chemical being used and the waste-water

8.3

8.4

portions into each of the test vessels.

period as shown in Table 2.

Place the stirrers in the vessels and start stirring at

concentration selected from results of Method A.

Stir the samples for the appropriate flash-mixing

the appropriate flash-mixing speed shown in Table 2 (note b). After 15 s add the chosen chemical to give a

Step	Procedure	Notes
8.5	Reduce the speed to the appropriate flocculation speed (shown in Table 2) and continue for the period shown (note c).	(c) If a second chemical, or coagulant aid is to be used, stir slowly for only 2 min, add the second chemical, stir at high speed for 30 s and proceed with step 8.5.

Table 2.

Step		Hydrated lime	Aluminium and iron salts	Polyelectrolytes	
Flash mixing (Coagulations)	Speed Period	100 rpm 2 min	100 rpm 2 min	300 rpm 30 s	
Flocculation	Speed Period	20 rpm 15 min	20 rpm 15 min	40 rpm 4 min	

Step	Procedure	Notes
8.6	Switch off and remove stirrers from the test vessels. Allow sample to settle for 1 hr (note d).	(d) A settling period of 15 or 30 min may be selected if using polyelectrolytes.
8.7	Measure temperature and pH value again (See 9.2).	
8.8	Drain off supernatant liquor via the upper (liquid) drain line in the base of the unit (note e).	(e) If supernatant liquor analysis as well as sludge quality and quantity is desired, withdraw samples from the middle of the supernatant liquor in volumes appropriate to the parameters to be determined.
8.9	When liquid drainage is complete drain off sludge via the lower pipe into a suitable container. Leave the sludge to settle for 1 h, and then decant off supernatant liquor. The volume and solids content of the thickened sludge should be estimated using the methods in this booklet. See also note f.	(f) Volume and solids concentration measured from further thickening by gravity in a shallow vessel will not normally represent the volumes achieved in a full-scale settlement tank but sludges produced in this way may be used to improve the measurement of CST (see 8.10).
8.10	Carry out additional sludge characterization such as Filtrability (note g), Specific Resistance to Filtration (SRF) (see method in this booklet) or Capillary Suction Time (CST) (see methods in this booklet).	(g) Filtrability is not as useful as a characteristic as the amount of additional chemical required to condition the sludge for satisfactory dewatering. Sufficient sludge for this measurement is not normally available from this test unless the procedure is repeated several times under identica conditions.
8.11	Repeat the procedure 8.1 to 8.10 for the next chemical of interest (note h).	(h) If different samples of the wastewater have been used in the determination of the optimum concentration for a number of chemicals, the test should be repeated by adding the optimum concentration of each chemical in step 8.3 to aliquot portions of a single sample of wastewater

Step	Procedure	Notes	Procedure	tes
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- 8.12 To compare chemicals, rank them in order of increasing volume, concentration, or other characteristic of the sludge produced (note i).
- (i) There is much confusion in the expression of concentrations of chemicals. For the purposes of comparison the concentrations should wherever possible be expressed as Al for aluminium salts, Fe for iron salts and Ca(OH)₂ for hydrated lime. Where the product has an indeterminate chemical composition, it is recommended that concentration be expressed as weight of a specified active ingredient per unit volume of liquid. For polyelectrolytes, and products which are mixtures of salts, concentrations should be expressed as weight of product per unit volume of liquid. For operational purposes it may be more convenient to express all concentrations as volume or weight of product supplied per unit volume of wastewater.

9. Sources of Error

- 9.1 Coagulation and flocculation may be strongly dependent upon pH value.
- 9.2 Results will be dependent on temperature. The test should be repeated if significant changes (more than 2-3°C) in temperature are observed. Continuous illumination of the samples may cause convection currents and disturb the settlement.
- 9.3 Some chemical solutions will lose their effectiveness with time, especially if made up with tap water. It is recommended that new solutions are made up regularly. It is also recommended that polyelectrolyte solutions at a concentration of 1% are prepared using de-ionized water. They can be stored for up to a week in the dark and diluted to 0.05% for the test.
- 9.4 The loss of carbon dioxide or other gases during the flocculation step may interfere with the chemical reaction and may lead to erroneous results.

10. Note on Floc Size and Floc Size Measurement (See Step 7.6 above)

Floc size is one of the important factors which determine the settleability and filterability of suspended solids. Precipitation is often necessary for the clarification of liquids, for the removal of undesirable substances from waters and effluents or for the concentration of materials from solution into the solid state. Material may itself be removed directly as the precipitate, engulfed in a precipitate or absorbed by one. The floc size of a precipitate is highly dependent on the conditions of preparation and on subsequent treatment. Factors affecting floc size are reagent concentration, the order and rate of mixing whether small seed crystals or particles are already present in suspension, if and when a coagulant or coagulant aid or filter aid is added, and whether a mixing process is used or flocculation occurs gradually from a relatively homogeneous solution.

Floc size and shapes are equally dependent on treatment after initial formation. Important factors in this category are violence of stirring, agitation, turbulence or handling which can break up flocs especially irregularly shaped ones; settling and thickening; washing and simmering or other variations in temperature including daily fluctuations. Washing can either increase or reduce floc size. If it removes adsorbed ions which are causing neutral particles, or particles with balanced positive and negative charges to repel each other, washing can lead to an increase in floc size, but if washing removes ions which have been attracted to opposite charges occurring in the surface of particles thus bonding them together, then washing can cause peptization. Temperature fluctuation as when particles circulate between the top and bottom of a simmering liquid or vary with room temperature causes alternating solution and redeposition which leads to the gradual disappearance of the smaller particles and the growth of the largest, it may lead to the slow crystallization of amorphous materials.

Flocs are often approximately spherical, but sausage shaped, tabular and dendritic forms are not uncommon. Likewise unless some form of natural or artificial size grading or conditioning has taken place variations in particle size are not uncommon and, it may be

advisable to grade giving the smallest and largest sizes noticed as well as the size of the majority, adding a short note if some or all particles are markedly not spheroidal, tend to float instead of sink or have other unusual tendencies. Because of this variability within one batch and the labile nature of most flocs very precise measurements are seldom required but some quantification is helpful in controlling sludge processes.

Two simple methods of estimating particle size are suggested, comparison with illustrations and inspection through grids of known size. Simple operator precision can often be achieved, but agreement between different operators can be poor, though the variations will rarely be serious. Comparative tests are recommended when changing the technicians doing this assessment.

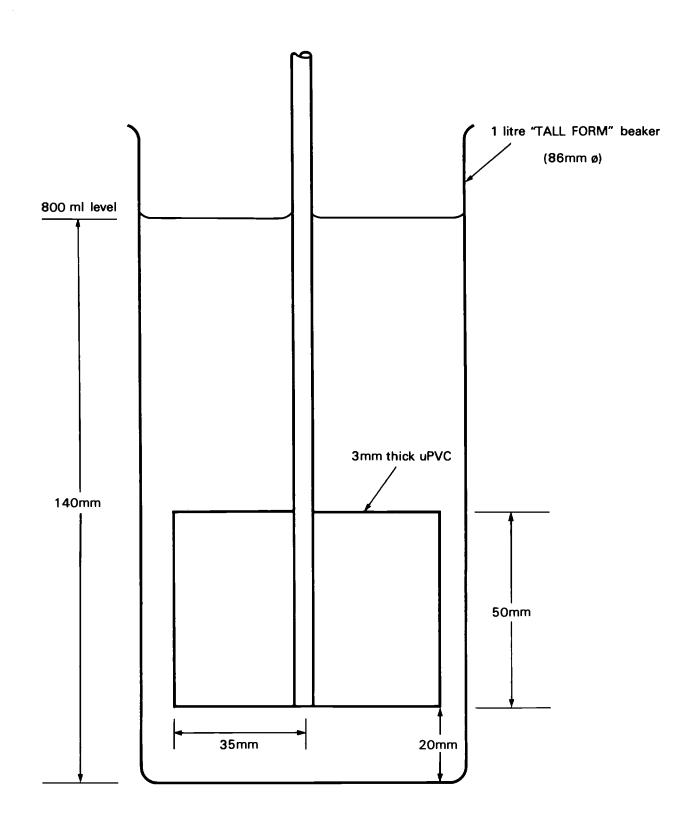


FIGURE 5 800ml TEST ARRANGEMENT

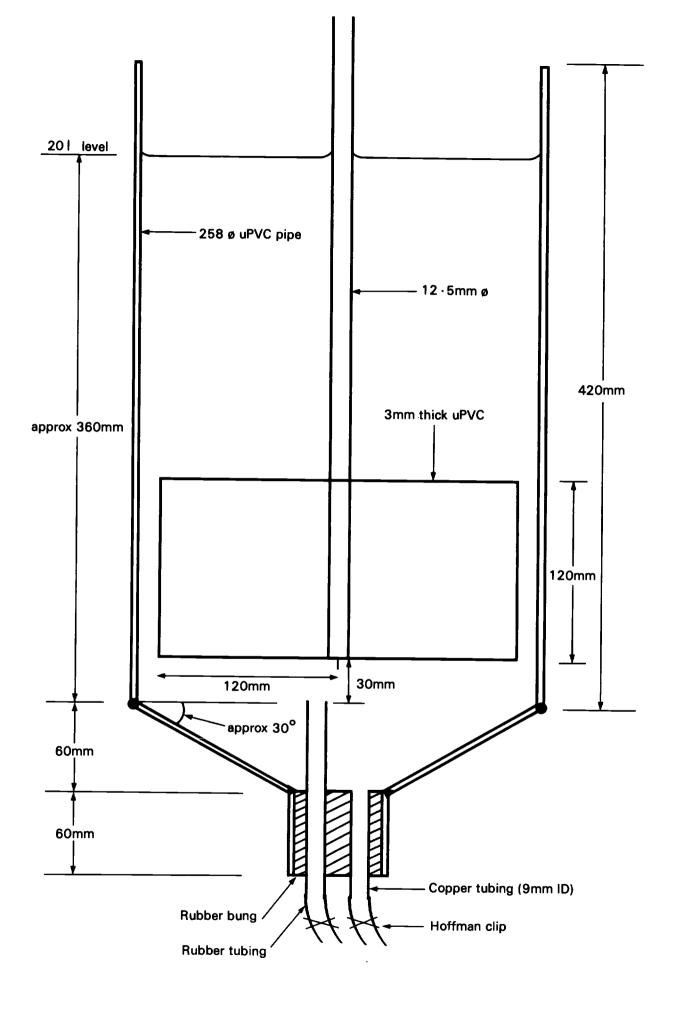


FIGURE 6 20 LITRE TEST ARRANGEMENT

Determination of the Thickening Characteristics of Sludges using a Low-Speed Centrifuge Method

1. Performance Characteristics of the Method

1.1	Parameters determined	 (a) Ultimate suspended solids concentration (C₀₀). (b) Compressibility index (B). (c) Thickening time (a). 	
1.2	Type of samples	Any suspension of solid particles having density greater than unity. Not suitable for very thin sludges where the concentration of suspended solids is less than 0.2% or those sludges that settle with more than one interface.	
1.3	Basis of the method	Observation of the fall of interface between the sludge and suprnatant liquor.	
1.4	Range of application	 (a) Ultimate suspended solids concentration usually within the range of 0.5-60% solids. (b) Compressibility indices between 0 and 0.4. Increasing index implies increasing compressibility. (c) Thickening time between 0 and 10³ min. Increasing number implies increasing time to achieve a particular solids concentration. 	
1.5	Standard deviation	Relative standard deviation usually better than (a) 3% (For C ₀₀) (b) 10% (For B) (c) 10% (For a)	
1.6	Interferences	None	
1.7	Time required for analysis	About 1 hour, but very dependent on sample.	

2. Principle (see also Ref. 5)

Samples of sludge of known initial suspended solids concentration (see method in this booklet) are centrifuged at 3 defined centrifugal accelerations and the fall in the interface level is observed with time, with the aid of a stroboscope synchronized to the speed of the centrifuge.

The various centrifugal accelerations used model different sludge blanket depths. The ultimate concentration can be determined either experimentally by observing the interface until there is no further movement or by extrapolating the results by calculation to infinite time.

The compressibility index is defined as the gradient of a logarithmic plot of the ultimate concentration against centrifugal acceleration.

The thickening time is defined as a coefficient in an empirical equation used to describe the change in suspended solids concentration with time.

3. Reagents

No reagents are used.

4. Apparatus

- 4.1 The low speed (frozen-image) centrifuge (Ref. 6)
- **4.2** Flat-bottomed graduated centrifuge tubes of diameter 14 and 26 mm, and depth 90 mm.
- 4.3 Two 250-ml glass beakers
- 4.4 100-ml widemouth sample bottle.

5. Sampling

It is important that the sample is truly representative of the sludge concerned (see publication in this series). Sludge samples may change composition through microbiological activity. It is important therefore that the samples be analysed as soon as possible. It is necessary to know the time elapsed from the time of taking of the samples so that the effect of storage on the thickening characteristics can be assessed.

6. Procedure

6.13

Repeat steps 6.2 to 6.12 using centrifuge speeds of

500 rpm and 300 rpm.

Step	Procedure	Notes
6.1	Select two appropriately sized centrifuge tubes (note a)	(a) The 26-mm diameter centrifuge tube should be used if the sludge is widely heterogeneous in particle size.
6.2	Fill tubes with water and place in carriers.	
6.3	Close centrifuge lid and start motor.	
6.4	Adjust speed of centrifuge to 700 rpm. (note b).	(b) Individual centrifuges can be precalibrated.
6.5	Stop centrifuge, leaving speed adjustment control at the 700 rpm position.	
6.6	Empty and dry centrifuge tubes.	
6.7	Thoroughly mix sample of sludge by pouring from one beaker to another.	
6.8	Pour sample into centrifuge tube up to 70-mm mark and place tube in carrier in centrifuge (note c).	(c) If two different samples are to be analysed, mix second sample as in 6.7 above and pour into the second centrifuge tube. If only one sample is to be analysed the second centrifuge tube can be used for a duplicate measurement.
6.9	Record placing of the different samples in the carriers, A and B and close centrifuge lid. (note d).	(d) Samples in positions A and B are observed separately by use of selector switch on control panel of centrifuge.
6.10	Start centrifuge and adjust speed, if necessary, to 700 rpm \pm 10 rpm.	
6.11	Record height of liquid in each centrifuge tube (note e).	(e) The initial height of sludge is used to calculate the change in suspended solids concentration with time.
6.12	Record the position of the interface at known time intervals (note f).	(f) Sludges thicken at highly variable rates and it is not possible to specify in advance an appropriate frequency for taking readings. Reading should be taken more frequently in the early stages of the test.

6.14 Calculate the average suspended concentration \overline{C}_t of the sludge in the centrifuge tube at time t for all the results using the equation

$$\overline{C}_{t} = \frac{\overline{C}_{0} h_{0}}{h_{t}}$$

where \overline{C}_o is the initial concentration in the centrifuge tube filled to a height h_o and h_t is the height of sludge from the base at time t (note g).

(g) This equation assumes that there are no suspended solids in the supernatant liquor. The error introduced by this assumption will in most cases be small; if, however, the supernatant liquor has a high concentration of suspended solids, C_sthen the equation

$$\overline{C}_{t} = \frac{\overline{C}_{o} h_{o}}{h_{t}} - \frac{\overline{C}_{s} (h_{o} - h_{t})}{h_{t}}$$

must be used.

7. Calculation of Parameters

7.1 Ultimate concentration

7.1.1 Direct method

Plot the results as suspended solids concentration against time for each of the three centrifugal accelerations. An example of such a graph is shown in Fig. 7. The ultimate concentrations may be determined from this graph where the tests can be continued to produce a plateau for each curve.

The ultimate concentration is dependent on the compressive force, which will be proportional to the product of the initial solids concentration, \overline{C}_o , and the centrifugal acceleration, \underline{a} , and so these must be stated when quoting the results.

7.1.2 Calculation

The equation

$$\overline{C} - \overline{C}_o = \frac{(\overline{C}_{00} - \overline{C}_o)(t - t_o)}{a + (t - t_o)}$$
 (1)

has been shown (Ref. 6) to fit the data from the centrifuge, where a is a time coefficient and t_o is the rearrangement time of the particles as illustrated in Fig. 8.

To determine the ultimate concentration, plot a graph of $\frac{t-t^*}{\overline{C}-\overline{C}_*}$ against $t-t^*$ where t^* is greater than t_o and \overline{C}^* the suspended solids concentration at t^* . See for example Fig. 9. The gradient of the straight line produced is equal to $1/\overline{C}_{00}$ - \overline{C}^*). The ultimate concentration, \overline{C}_{00} is calculated from the equation;

$$\overline{C}_{\infty} = \frac{(1)}{\text{Gradient}} + \overline{C}^*$$
 (2)

Thus in Fig. 9,
$$C_{\infty} = \frac{1}{0.67} + 1.45$$
 at = 2.9% (C_o = 0.55, a = 45)

7.2 Compressibility index

Having calculated the ultimate concentration for each of the centrifugal accelerations using either of the two methods above, plot a logarithmic graph of the ultimate concentration (y-axis) against centrifugal acceleration (x-axis), as shown in Fig. 10. The compressibility index is the gradient of this line.

7.3 Thickening time

The intercept on the graph of $\frac{t-t^*}{C-C_*}$ against $t-t^*$ (Figure 9) is used to calculate the thickening time. The time coefficient a from equation 1 is defined as the thickening time and is calculated from the equation

$$a = Intercept \frac{(\overline{C}_{00} - \overline{C}^*)^2}{\overline{C}_{00} - \overline{C}_0}$$
 (3)

The thickening time is dependent on the centrifugal acceleration and the initial concentration. The results should be quoted at 45 g and the initial solids concentration clearly stated. For example; the thickening time for the results shown in Fig. 7 is 1.5 min from an initial solids concentration of 0.55% and written as a = 1.5 min ($C_0 = 0.55\%$).

The value of to may be calculated from the equation

$$t_o = t^* - Intercept \frac{(\overline{C}^* - \overline{C}_o)(\overline{C} - \overline{C}^*)}{(C - C_o)}$$
(4)

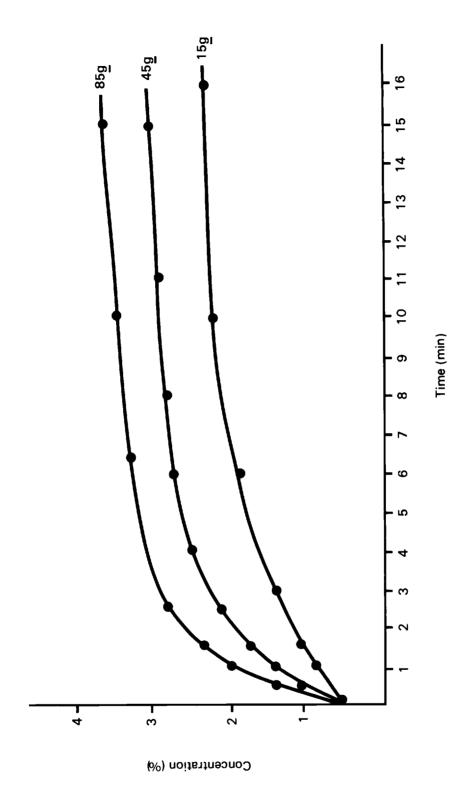
Interpretation of Results

The centrifuge speeds of 700, 500 and 300 rpm produce centrifugal accelerations at the base of the centrifuge tubes of 85, 45 and 15 g respectively. The heights of sludge in a thickener (H), in metres, to which these centrifugal accelerations correspond are calculated using the equation

$$H = 0.07 \underline{a} \tag{5}$$

where \underline{a} is the centrifugal acceleration, provided the initial suspended solids concentrations are the same in the thickener and centrifuge.

The thickening time, a, is approximately the time required in the centrifuge for the sludge to achieve the concentration half-way between the initial and ultimate concentrations, (ie. $\bar{C}_o + (\bar{C}_{00} - \bar{C}_o)/2$). If the value of a is small compared to to then the time to reach the half-way concentration is $(a + t_o)$.



AT A NUMBER OF CENTRIFUGAL ACCELERATIONS FOR A SURPLUS ACTIVATED SLUDGE SUSPENDED SOLIDS CONCENTRATION PLOTTED AGAINST TIME IN THE CENTRIFUGE FIGURE 7

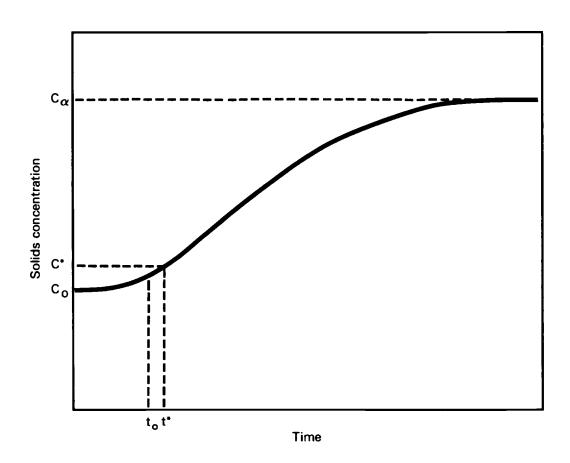
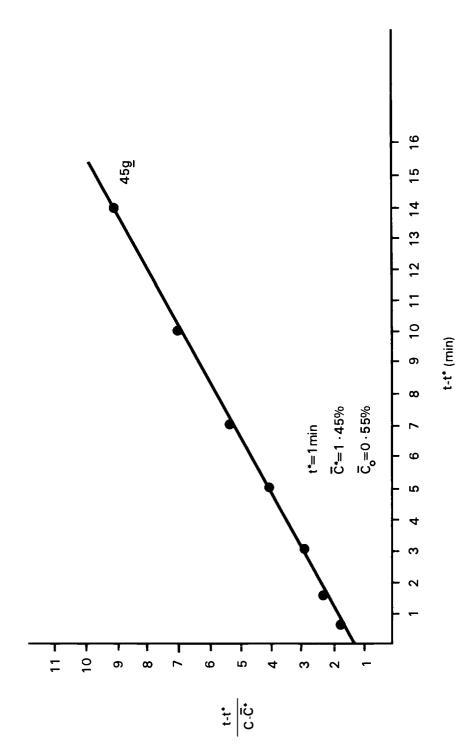


FIGURE 8 A TYPICAL THICKENING CURVE



SHOWING LINEARITY OBTAINED BY PLOTTING t-t*/C-C* AGAINST t-t*. FIGURE 9

Determination of the Settleability of Activated Sludge Expressed as the Stirred Specific Volume Index (SSVI) or Stirred Sludge Density (SSD)

Performance Characteristics of the Methods

1.1	Parameters determined	(a) Stirred specific volume index (SSVI).(b) Stirred sludge density (SSD).
1.2	Type of sample	Activated sludge.
1.3	Basis of method	Settlement of sludge of a known solids concentration in a stirred cylinder for 0.5 hour
1.4	Range of application	Depends on solids content of sludge (see Section 8)
1.5	Standard deviation	Sample dependent but normally less than 10% Relative Standard Deviation, except for sludges with very long settling times, and sludges that will not settle at all.
1.6	Interferences	See Section 3.
1.7	Time required for analysis	0.75 h per settlement test. Overall time approximately 3 h including determination of sludge solids (SS) (see method in this booklet).

2. Principle

The volume of sludge after 0.5 h settlement in a standard stirred cylinder is measured, and a specific volume is derived by dividing this figure by the weight of sludge solids present, determined gravimetrically by the method described in this booklet.

3. Interferences

Vibration or rapid changes in temperature may influence the result obtained.

4. Reagents

No reagents are required.

5. Apparatus

The recommended apparatus consists of a polymethylmethacrylate cylinder, 10-cm diameter, marked with a 50-cm vertical scale, and fitted with a 1 rev/min stirrer, as shown in Fig. 11.

Stop clock.

Apparatus for determination of suspended solids.

6. Sampling

It is most important that the sample is truly representative of the sludge concerned (see publication in this series Ref 10). Sludge samples may change composition through microbiological activity and it is important that the settlement test is carried out immediately.

7. Procedure

Procedure

Step

one vessel to another several times to make sure the sample is homogeneous. Immediately take a small subsu			
and quickly pour the rest of the sample into a stirred settlement apparatus (note b) approximately up to the 50-cm mark (note c). Connect up the stirrer, start the timing device and record the height of the suspension su	7.1	mes to make sure the iately take a small subpended solids (note a) esample into a stirred oproximately up to the property the stirrer, start the ight of the suspension (b) The suspension (c) If the	ne a rfa exc)-cn onta spe spe

7.2 After 0.5 h record the height of the suspension again. Let this height be h₁ cm.

Calculation

7.3 Let the initial concentration of suspended solids be C_o per cent w/w

(a) SSVI =
$$\frac{100 \text{ h}_{1}}{\text{C}_{0} \text{ h}_{0}} \text{ ml/g}$$

(b) SSD =
$$\frac{C_{\circ} h_{\circ}}{h_{\perp}}$$
 per cent

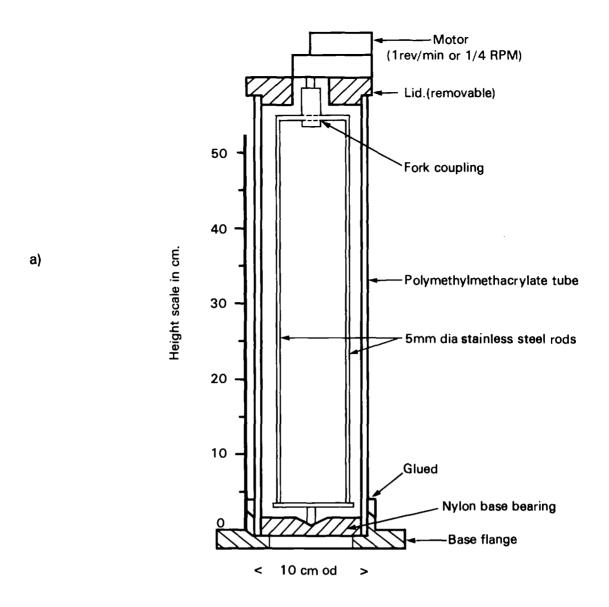
Notes

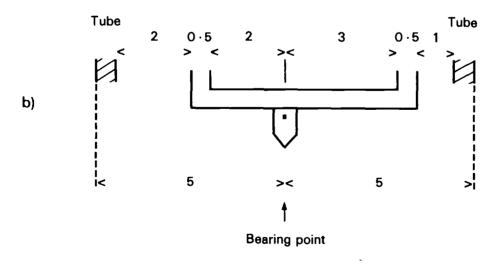
- (a) See method published in this booklet.
- (b) The apparatus should be on a level horizontal surface.
- (c) If excessive care is taken to fill the apparatus to the 50-cm mark the final liquid poured in may not contain a representative concentration of suspended solids and hence the concentration of suspended solids determined in the sub-sample will not relate to that in the apparatus.

8. Effect of Solids Content

The parameters SSVI and SSD depend on the solids content of the sludge. Solids contents of activated sludge are usually in the range 0.5 to 10 g/l. For a typical sludge of 3.5 g/l, for example, the SSVI will not exceed 280 ml/g. Correspondingly, SSD will be not less than 0.35 per cent (when the sludge does not settle) and will probably not exceed 6.7 per cent.

Determination of SSVI or SSD at a solid concentration of 3.5 g/l enables the solids loading on secondary settlement tanks to be estimated (Refs. 8,9).





Approximate cross-sectional dimensions of tube and thickener rods in cms

FIGURE 11 STANDARD SETTLING APPARATUS

Measurement of the Specific Resistance to Dewatering of a Sludge and Sludge Cake Compressibility

Introduction

Although solids can often be separated from liquids by settling or by flotation, filtration is usually necessary if only to polish the effluent and is usually quicker though more expensive. Similarly, although filtration can be achieved by diffusion of the liquid away from the solids when placed on the surface of a filter medium, practical filtration requires the use of some form of pressure which can be simply the force of gravity, but industrially is usually either suction as in a vacuum filter, applied pressure to the mixture as in a filter press, or an overall force as in centrifugal filters. There are many variations in the design of filters but the basic principle is always the same. A suspension of a solid in a liquid passes through a medium with pores smaller than the size of the particles to be retained, under an applied pressure gradient. As filtration proceeds a cake builds up on the filter which itself acts as a filter. If there are fine particles or compactable particles in the suspension they may clog either the filter medium or the pores in the filter cake thus blinding the filter and eventually stopping filtration. Some materials undergo some form of re-structuring during filtration resulting in a denser more compact cake. Metal oxides and hydroxides often do this by chemical condensation or recrystallization. Even so, many filter cakes contain as much as 70-80% of retained moisture and over 90% is not unknown. Various filter aids and conditions are used to improve the filterability of sludges so as to achieve the maximum economic effluent removal in minimum time with as little filter blinding as possible. Similarly, squeezing the filter cake, by increasing the filtration pressure gradient, can produce a denser cake, though at the expense of reducing pore size and causing filtration to slow down.

Over the years, two parameters have been devised for measuring the ease of dewatering a filter cake and of assessing its response to pressure changes.

These are (1) The Specific Resistance, and

(2) the Cake Compressibility.

1.1 Specific Resistance

The time taken to filter a given volume of liquid is given by the formula

$$\frac{t}{V} = \frac{u r C}{2PA^2} V = \frac{uR}{PA}$$

where t is the time taken to filter the volume V,

u is the filtrate viscosity,

r is the specific resistance,C is the sludge concentration,

P is the pressure difference across the filter, including cake,

A is the filter area, and

R is the filter medium resistance.

The filtration rate is dependent on two factors: the resistance of the cake, given by $\frac{\text{urC}}{2PA^2}$ V and the resistance of the filter medium, which includes the effect of blinding of the pores in the filter medium itself, and which can change with time as the filter medium gets older.

The Specific Resistance is usually measured by plotting the time per unit volume (t/V) against V keeping all other variables constant, which includes the temperature, not included in the above equation but which is a variable affecting the filtrate viscosity.

The gradient of this plot of t/V against V

is usually written b which equals $\frac{urC}{2PA^2}$

Hence
$$r = \frac{b2PA^2}{uC}$$

The specific resistance of a cake can therefore be measured for any constant set of conditions. Knowing the specific resistance it is then also possible, if desired to calculate the filter medium resistance (R) to filtration of that sludge.

1.2 Cake Compressibility

As mentioned above the specific resistance (r) varies with the applied pressure gradient across the cake. This variation is usually measured by the coefficient of compressibility (S) which is given by $r = r_0 PS$

where r is the specific resistance at pressure P and r_o is a constant.

Hence if a series of tests is made in which the specific resistance is measured at different pressures, all other variables being kept constant and a log/log plot made of the specific resistance against pressure, the gradient of this plot is the cake compressibility coefficient (S).

1.3 Units

The units used have varied over the years. At the present time the following units are used for the measurements

Time t in seconds (S)
Volume V in metre cube

Viscosity N in Newton seconds per square metre (Ns/m²)

(see below)

Sludge Concentration C in kilograms per metre cube (kg/m³)

Pressure P in Pascals (Pa)
Filter area A in square metres

Giving the Specific Resistance r in metres per kilogram

The filter medium resistance is measured in reciprocal metres (m⁻¹).

The cake compressibility coefficient S is a pure number.

Measurement of the sludge concentration is often taken as the total solids content, but if the dissolved solids content is high, the suspended solids value would be more appropriate. Similarly the viscosity of the effluent filtrate is usually assumed to be the same as for water, but this may not necessarily be true. Information on measuring viscosity, which in turn entails measuring the density is given with the method.

For further information see Refs. 11 and 12.

2. Interpretation

From consideration of the equations for rate of filtration above, the smaller the value of the Specific Resistance and of the Filter Medium Resistance the more rapid the filtration should be.

3. Performance Characteristics of the Method

3.1	Parameters determined	(a) Specific Resistance to Dewatering(b) Sludge Cake Compressibility(c) Filter Medium Resistance
3.2	Type of Sample and Problem Investigated	Sewage, Waterworks and similar sludges and slurries and also evaluation of filter aids, and coagulants for given sludges. Also the effect of pressure and temperature on plant operation.

3.3	Basis of the method	Measurement of the variation in the filtration rate with time at various pressure drops at constant temperature.
3.4	Reliability and Application	Measurements are usually for qualitative use to determine if one set of conditions or coagulant procedure is better than another and the effect of changing pressure. The biggest source of variability in obtaining the measurements is sludge sample inhomogeneity; results should not vary by more than 10% of the value determined.
3.5	Time required	This is highly dependent on the way in which the tests are carried out and the equipment used; but is usually several days for a complete study.

4. Principle

An appropriate filtration apparatus is assembled with either the filter medium to be used or some suitable substitute. The sludge-effluent sample to be evaluated is treated with any initial pretreatment that may be necessary, and then filtered at constant pressure and temperature, the volume of filtrate produced being measured for a convenient length of time. If the effect of pressure variation is being studied the test is repeated several more times with different overall pressure differences across the filter the temperature remaining unchanged. Similarly if the effect of temperature is being investigated the test would be repeated several times more with the same pressure drops but with variation in the temperature.

5. Hazards

Apparatus at above or below atmospheric pressure needs suitable guards to protect operators. Glass and ceramic vacuum filtration apparatus must always be surrounded by a guard in case of implosion. Unless relatively high pressures are used the greatest risk from apparatus at elevated pressures is from leaks spurting. Adequate blast walls or screens should be used if there is any risk of pressure vessels bursting. Adequate eye protection must be worn (visor and safety glasses which do not protect against liquid which might spray round their sides are not suitable). Compressed gas and Pressure Vessel Regulations should be consulted and obeyed.

6. Apparatus

- 6.1 Filter this will be dependent on the type of filtration being studied, and should be chosen keeping in mind the full scale plant to which the results will be applied.
- 6.1.1 Vacuum a large filter flask, connected to a source of vacuum by a regulating valve (or an alternative device for obtaining a steady sub-atmospheric pressure may be used), fitted with a vacuum pressure gauge or manometer, and with a deep Buchner type filter funnel. The filter paper or cloth used must sit on a fine grid or nylon gauze so as to ensure that the whole surface is used at the same pressure. A narrow rubber or similar collar should fit tightly inside the funnel, pressing down onto the filter paper or cloth to prevent channelling round the edges of the filter medium at the filter funnel walls.
- 6.1.2 Pressure Buchner Funnel (a suitable apparatus is illustrated in ref 11). A metal funnel is assembled with screwed or flanged joints and rubber gaskets holding a perforated filter support plate, a distribution grid and the filter paper or cloth to resemble a vacuum filter funnel as in 6.1.1 above; but a tubular metal body screws down onto the filter to form the sludge slurry reservoir. The top of this reservoir is closed by a gasketted cap firmly secured, and either the cap or top of the reservoir carries a compressed air connection to which is also fitted a pressure regulating valve, a pressure gauge and a vent valve. A master pressure isolation valve must be installed within easy reach of the operator, between the air supply line and the apparatus.
- 6.1.3 Pressure plate press small commercially available plate and frame presses can be assembled with only one frame unit and used for these measurements. Leaf and other similar bag filters are not recommended for test purposes.
- 6.1.4 The tests can also be made on full size plant provided suitably calibrated tanks for measuring the effluent delivered in various times (or amount of sludge used) are available and the various other measurements are obtainable with sufficient accuracy.

6.3 Measures

- 6.4 Other equipment may be needed in order to make other measurements needed in the calculation. See Procedure steps 8.2.1, 8.2.2, 8.3, 8.7 and 13.
- 6.5 Provision should also be made for keeping the temperature constant during the tests ($\pm 2^{\circ}$ C).

7. Sample Preparation

Prepare a suitable stock of slurry for each test, with any necessary conditioner, filter aid or necessary flocculation, ensuring as far as possible that the overall within batch and between batch composition is as constant as possible.

B. Test Procedure

All Calculations

It is impossible to lay down one single procedure, as what is done is dependent on the equipment used, why the test is being carried out, and the type of filtration being studied. However basic guidelines are possible. The procedure must be kept standardized for each type of test a laboratory makes.

- 8.1 The apparatus is assembled in such a way that the volume of filtrate can be measured after known intervals of time by collection directly either in a measuring cylinder or in another suitably calibrated vessel. (A previously weighed container mounted on accurate heavy load scales for recording increases in weight during the test may be useful in some instances provided the filtrate density is determined or known.)
- 8.1.1 If this is not possible a few other alternative ways of measuring the filtrate flow rate may be possible, such as by being able to stop filtration without damaging the filter cake or pressure seals (this may be difficult or impossible), decant the filtrate for measurement and then resume filtration, or by running batch filtrations with new filters filtering progressively larger amounts of slurry and timing the filtration.
- 8.1.2 Another alternative which may be possible for thin slurries is to measure the slurry input. This is only usable provided it is possible to measure in separate small-scale laboratory tests the amount of filtrate obtainable from a sample of the same slurry, in a filtration giving a cake of the same solids content as the sludges obtained in the main tests.
- 8.2 Using a separate representative sample of the slurry to be filtered determine either the total solids content or the suspended solids content, whichever is more appropriate (see Section 1.3 above).

8.2.1 Determination of Total Solids Content

See the first set of methods in this compendium.

8.2.2 Determination of the Suspended Solids Content

See the second set of methods in this compendium, or the corresponding method in reference 1. Choose the appropriate method of filtration.

8.3 Measure the area of filter to be covered with filter cake. Allow for areas covered by gaskets or otherwise obscured and not available for filtration.

If calculation from direct measurements is too difficult the area can be measured by marking the filter size, allowing for gaskets etc, onto a sheet of 1mm squared graph paper and counting squares (such papers always have centimetre and larger squares also indicated).

8.4 If the filter paper or cloth needs pre-coating with a filter aid in order to prevent blinding or in order to aid release of the cake after filtration, do that prior to starting slurry filtration. Such pre-coating should be carried out in a reproducible manner and all old coating removed and replaced before replicate runs are made.

8.5 Add the sample slurry to the apparatus and filter at constant known pressure and constant known temperature, noting the time taken to collect known amounts of filtrate (or to filter known amounts of slurry of known filtrate content, see 8.1.2 above).

Take sufficient volume-time readings to ensure that a reliable gradient can be measured for the plot of time/volume versus volume. Readings should preferably be over a tenfold to one hundredfold range in volumes.

- 8.5.1 If a centrifugal filter is used keep the rotation rate and drum charge depths constant.
- 8.5.2 If the initial filtrate is cloudy but subsequently becomes bright, note both the time and filtrate volume at which this occurs. There is a possibility that the graph may not be linear before this point as the quality of filtration has been changing. (See Sections 9.1.1 and 12.4 below.)

Sludge Cake Compression Coefficient

8.6 Repeat the above test at several different filtration pressures by adjusting either the vacuum, or applied inlet pressures (or for centrifugal filters, centrifuge speed).

Effect of Temperature

8.7 If it is desired to study the effect of temperature, arrange for both filter and incoming slurry temperatures to be adjusted to the temperature desired and kept as constant as possible throughout while the test is repeated as often as necessary. Record the temperature of the slurry just prior to the filter during each test.

9. Calculation of the Specific Resistance to Dewatering

Ensure that all units used for measurement are converted to accord with Section 1.3.

- 9.1 From the test data calculate the various values of
- t_x/V_x where t_x is the time taken to collect volume V_x .

Plot the graph of t_x/V_x against V_x .

9.1.1 If the filtrate took a significant length of time (t_c) to become clear (step 8.5.1 above) and the graph is not vertically linear, calculate

$$\frac{t_x - t_c}{V_x - V_c}$$
 and

plot
$$\frac{t_x - t_c}{V_x - V_c}$$
 against $V_x - V_c$

- 9.2 Measure the gradient of the graph in 9.1 (or 9.1.1) let this be b.
- 9.3 The Specific Resistance is

where b is as in 9.2

- P is the constant filtration pressure drop* (See footnote on p 50)
- A is the available filter area (steps 8.3)
- u is the filtrate viscosity. If markedly different from that of water or if it is desirable to check this see Section 13.
- C is the solids concentration of the slurry (step 8.2).

10. Calculation of the Sludge Cake Compressibility Coefficient

10.1 Determine the Specific Resistance to Dewatering(r) (Section 9) for a range of pressure drops(P).

10.2 Plot the logarithm of specific resistance against the logarithm of pressure drop (log r versus log P) and measure the gradient of the curve (S): this is the sludge cake compressibility coefficient.

11. Calculation of the Filter Medium Resistance

From Section 1.1

$$R = \frac{PAt}{uV} - \frac{rCV}{2A}$$

Using data obtained in Sections 8 and 9 substitute in the above formula to obtain R.

12. Sources of Error

The following factors are important sources of error.

12.1 Reproducibility of the Sludge Slurry composition.

12.2 Constancy of pressure drop and temperature.

12.3 Reproducibility of filter medium

This is especially true if filter cloths are used and more so if filter aid coated cloth is used.

12.4 Variation in effective filter pore size with time

If there is a wide variation in the sizes of particles being filtered, at the start of filtration only large particles may be retained, small ones being passed, but the large particles then constitute a filter which may retain the small particles, in which the t/V versus V curve may not be linear until the filtrate becomes clear. (See Section 8.5.2 and 9.11 above).

12.5 Approximations for sludge solids concentration and filtrate viscosity (see Section 13 and 12.8 below).

12.6 For centrifugal drum filters, the pressure calculation which is based on

Pressure = Force/unit area

= load mass/unit area \times (angular velocity in radians)² \times radius.

It is likely that the drum loading will be uniform over the surface from top to bottom. The solids content will contribute to the pressure on the drum itself, but may not contribute directly to the fluid pressure. (In the calculation in section 9.3 above only the liquid density is considered.)

12.7 For pressure and vacuum filters there will also be a contribution to the pressure from the head of liquid in the filter, usually small and dependent on whether the filter is vertical or horizontal. For plate and frame presses this will give a small gradient down the filter.

12.8 Other small inaccuracies of measurement such as loss of head (pressure) due to the kinetic energy of liquids (especially in calculating viscosity where measurement should be so regulated that drops drip and not flow, to avoid this error). Correction is possible but complicated see Ref 13. Another similar minor error is due to curvatures of meniscus. However, in the measurement of the capillary bore used for absolute viscosity measurement make the length measured many times the bore to minimize error due to the meniscus the measured length would be for a true cylinder of that amount of mercury.

For centrifuges $P = \frac{dl(2V)^2a}{3600}$

where d is the density of the slurry,

l is the cake thickness on the wall

V is the number of revolutions per minute

and a is the internal radius of the drum

^{*} For vacuum filters deduct the vacuum pressure from atmospheric pressure

13. Measurement of Filtrate Viscosity

Filtrate viscosity is often assumed to be that of water which varies between 0.018 and 0.010 c.g.s units at 0°C and 20°C respectively. This assumption may not necessarily be valid for moderately strong solutions.

The usual methods for measuring viscosity are by measurement of flow through a capillary tube under constant pressure at constant temperature, or comparatively either by comparing the time taken for liquids to flow through a viscosimeter or by comparing the time taken for a very small ball to fall under standard conditions in both liquids (see Ref 13 or similar text for further information on all methods).

For some of these measurements a knowledge of the effluent density may be required. For strong solutions, this too may differ significantly from that for water and should be checked.

14. Analytical Quality Control

As mentioned in Section 12, the biggest source of error in this test is variability of the material filtered. Sludge suspensions tend to change characteristics on storage, hence if routine laboratory checks on reproducibility are to be made recourse must be made to a freshly prepared synthetic material such as ferric or aluminium hydroxides prepared to a carefully controlled recipe and used at once as even such materials can vary with method of preparation and age with time.

On the other hand, if due allowance is made for variation in the sample, a control chart, with investigation of deviations, may be used for control purposes, not only of the analytical procedure but of the plant operation itself. (See Refs 14 and 15.)

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Address for Correspondence

Users of these methods are asked to send their comments and relevant data derived from use of the methods to the address below in order that the methods can be revised and improved in the light of experience.

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